

ESTIMATING BIOMASS OF *MELALEUCA CAJUPUTI* FOREST ECOSYSTEMS BASED ON MULTI-SOURCE SATELLITE DATA IN MEKONG DELTA, VIETNAM

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Abstract. *Melaleuca* forests in the U Minh Thuong ASEAN Heritage Park, the Mekong Delta of Vietnam involve a distinctive ecosystem that contributes significantly to biodiversity conservation and the provision of essential ecosystem services. Despite their importance, methods for accurately estimating carbon stocks in these forests remain insufficiently studied. This study utilizes both RADAR and optical data from ALOS-2, Sentinel-1, Landsat 8, and Sentinel-2 satellites to model and map the spatial distribution of aboveground biomass (AGB) in *Melaleuca* forests within U Minh Thuong National Park, a tropical forest site in the Mekong Delta of Vietnam. Various models were constructed using the backscatter values and Gray Level Co-occurrence Matrix (GLCM) calculated from RADAR data and normalized difference vegetation index (NDVI) from optical data, and field estimated AGB. For modelling, k-fold cross-validation approach was used. The study also found that the model integrating ALOS-2 PALSAR-2 HV data, GLCM textures, and Sentinel-2 NDVI, showed the best results ($R^2 = 0.84$, RMSE = 9.59 Mg ha⁻¹). The research underscores the significance of remote sensing data in supporting conservation efforts and sustainable forest management practices in tropical ecosystems. The study contributes to advancing biomass estimation methodologies and emphasizes the importance of multi-sensor satellite data integration for accurate and reliable forest biomass mapping, particularly in ecologically sensitive regions like the U Minh Thuong National Park.

Keywords: *aboveground biomass, U Minh Thuong National Park, Melaleuca forest, heritage park, ALOS-2 PALSAR-2, Sentinel 1&2, Landsat 8 OLI*

Introduction

Melaleuca cajuputi forests form a unique ecosystem in the coastal wetlands in tropical and subtropical zones (Tran et al., 2013). They are primarily located in Australia, New Guinea, the southern parts of United States, the Caribbean, Malaysia, Thailand, Cambodia, and Vietnam (Bartolo, 2005; Tran et al., 2013). In Vietnam, they are concentrated in the Mekong Delta. These forests provide vital ecological and socio-economic services, supporting biodiversity by offering habitats for various species and supplying timber, charcoal, fuelwood, oil, and honey. The carbon sequestration of these forests plays a critical role in mitigating climate change (Bartolo et al., 2005; Tran et al., 2005; RSIS, 2015; Xiang et al., 2025). Therefore, precise biomass estimation is essential for evaluating carbon stock levels (Raihan et al., 2021; Huang et al., 2022).

Traditional aboveground biomass (AGB) assessment methods rely on extensive field inventories, which are labour-intensive and impractical in inaccessible areas, limiting their use to small regions. In contrast, geospatial technology provides a more efficient and cost-effective approach for AGB mapping, monitoring, and modelling in large regions

(Lu, 2006; Kushwaha et al., 2014; Manna et al., 2014; Kalita et al., 2022). However, field data continue to play a crucial role in the training and validation of remote sensing (RS)-based AGB estimates. An extensive array of RS datasets, including optical multispectral and hyperspectral imagery as well as active sensor data such as light detection and ranging (LiDAR) and synthetic aperture radar (SAR), are now readily accessible and increasingly utilized for AGB estimation and monitoring. These datasets offer varying spatial, spectral, and temporal resolutions, enabling researchers to capture vegetation structure, canopy properties, and biomass-related parameters with improved accuracy and scale (Li et al., 2018; Nandy et al., 2019; Singh et al., 2023; Luong et al., 2025).

Recent technological advancements and the growing availability of data have allowed researchers to combine various satellite datasets, enhancing the precision of forest biomass estimation (Luong et al., 2019; Dang et al., 2019; Opelele et al., 2021; Bhandari and Nandy, 2024). However, challenges remain, particularly in large, complex environments such as tropical forests. Additionally, the lack of equations for converting biomass from the basic parameters of individual forest trees has led to discrepancies in research findings, undermining their scientific robustness (Baskent et al., 1995; Xiao et al., 2019; Huang et al., 2022). SAR remote sensing is ideally suited for tropical regions, as it penetrates cloud cover and dense vegetation. Its all-weather, day-and-night capability enables reliable estimation of forest structure as well as biomass (Mougin et al., 1999; Hamdan et al., 2015; Luong et al., 2016; Omar et al., 2017; Tuong et al., 2019; Golshani et al., 2019; Luong et al., 2024).

Despite early efforts to use remote sensing data for mapping *Melaleuca* forests (Williams, 1984; Bartolo, 2005; Ibrahim et al., 2012), there has been a lack of attention towards estimating *Melaleuca* forest biomass using remote sensing, particularly in Vietnam (Bartolo, 2005). Therefore, this study aims to develop models, using various combinations of ALOS-2 PALSAR-2, Sentinel-1, Sentinel-2 and Landsat 8 OLI, for AGB estimation and to map the AGB in *Melaleuca* forests in U Minh Thuong National Park, Mekong Delta in Vietnam. This will support conservation efforts and promote sustainable forest management within this unique tropical ecosystem.

Method and material

Study area

The U Minh Thuong National Park is located to the westward of the Ca Mau peninsula in An Giang Province, Mekong Delta in Vietnam. It is approximately 365 km southwest of Ho Chi Minh City. The park covers a total area of 8,038 ha (Tran, 2013). The study area (9°31'16" to 9°39'45" N latitude and 105°03'06" to 105°07'59" E longitude) is geographically located in An Minh Bac and Minh Thuan communes of U Minh Thuong District, An Giang Province (*Figure 1*). Importantly, the park is designated as one of the top three sites prioritized for wetland conservation in Vietnam's Mekong Delta region (RSIS, 2015).

U Minh Thuong National Park harbours a distinctive ecosystem characterized by *Melaleuca* forests on swam and peat. The dominant arboreal species, *Melaleuca* (*Melaleuca cajuputi*), forms the predominant constituent of the forest canopy. However, alongside *Melaleuca*, several other tree species such as *Alstonia spathulata*, *Ilex cymosa*, *Euodia lepta*, *Syzygium cumini*, *Elaeocarpus hygrophylus*, as well as various grasses and shrubs including *Phragmites vallatoria*, *Eleocharis dulcis*, *Stenochlaena palustris* and *Nephrolepis falcate*, are also found in the forest. The park supports numerous rare and

threatened species. The rare plant species recorded include *Elaeocarpus hygrophilus* Kurz., *Hydnophytum formicarum* Jack., *Eulophia graminea* Lindl., *Pachystoma pubescens* Blume., and *Spiranthes sinensis* (Pers.) Ames. The rare animal species recorded include *Lutra sumatrana*, *Prionailurus viverrinus*, *Manis javanica*, *Pteropus lylei*, *Leptoptilos javanicus*, *Mycteria leucocephala*, *Pelecanus philippensis*, *Cuora amboinensis*, *Heosemys annandalii*, and *Malayemys subtrijuga*, and its *Melaleuca* forest on swamp peatland represents a distinctive and endemic ecosystem of the lower Mekong Delta in Vietnam (Tran, 2013; Truong et al., 2024).

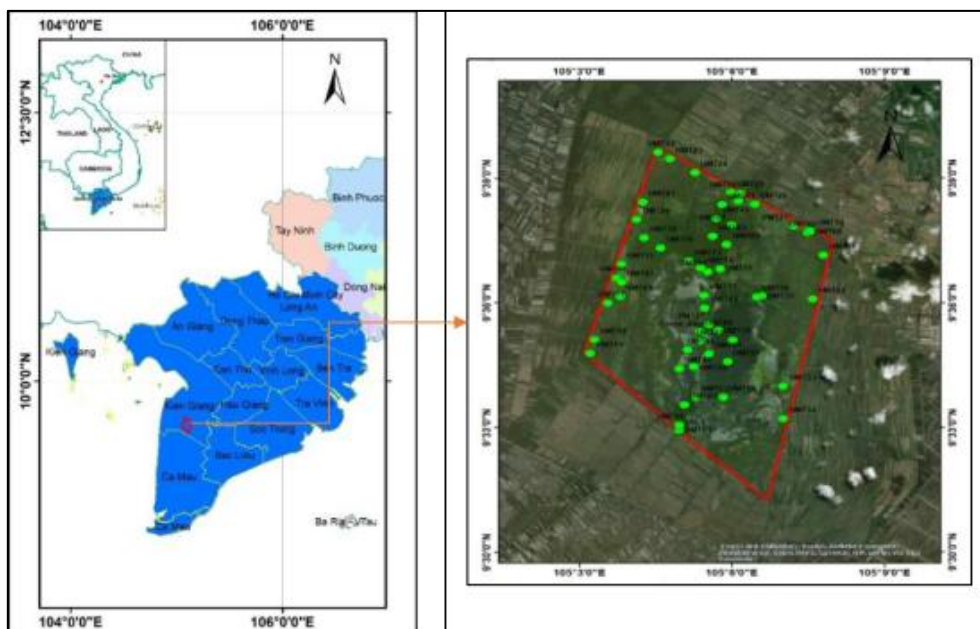


Figure 1. Study area map with sample plot locations represented by green points

The park, recognized in August 2013 as the first ASEAN Heritage Park designated on peatland in Vietnam, plays a crucial role in conserving its extensive peat layers and acting as a natural hydrological buffer by regulating groundwater and distributing surface water to nearby areas, essential for sustaining local livelihoods. The park also holds significant spiritual, historical, archaeological, educational, and scientific value (Tsitsi, 2016).

Some field survey photos are shown in *Figure 2*.

Methodology

Field data collection and biomass calculation

Sample plots were established across the study area during a field survey carried out in October 2018. The size of each sample plot is 500 m² (20 m × 25 m). A total of 54 sample plots were established and evenly distributed across the study area. The location and distribution of sample plots are shown in *Figure 1*. In each sample plots, the diameter at 1.3 m position, and height of all trees having a diameter greater than 5 cm were measured using laser measured by diameter laser meters (Criterion RD1000) and lasers measuring height (Trupulse 360B). The device was manufactured by Laser Technology Inc. (LTI), Origin: USA. The geographic coordinates at the centre of all the sample plots

were recorded using Garmin GPSMAP 87S GPS instrument (Garmin Ltd., Origin: USA). AGB was calculated using *Equation 1* (Dan et al., 2014).

$$AGB = 0.3188 \times D_{1.3}^{2.1672} \quad (\text{Eq.1})$$

where, *AGB* is the aboveground biomass of a tree (kg); $D_{1.3}$ is the diameter at 1.3 m position.



Figure 2. Field survey photos of the study area

The calculated fresh biomass was then converted to dry biomass by multiplying with the conversion factor of 0.5, following Hồng et al. (2015). A summary of the field survey results is presented in *Table 1*.

Table 1. Summary of the surveyed variables in the research site

No.	Forest variable	Minimum	Maximum	Median	Standard deviation
1	Diameter (cm)	5.61	23.16	8.44	2.74
2	Height (m)	3.66	12.16	5.14	2.05
3	Density (tree ha ⁻¹)	160	6260	550	1,550
4	Biomass (Mg ha ⁻¹)	2.20	115.30	10.12	27.21

Satellite data

The satellite data used in the study was selected based upon the accessibility, present and future availability, and their applicability in biomass studies. Synthetic Aperture Radar (SAR) data acquired from ALOS-2 PALSAR-2 (L-band) and Sentinel-1 (C-band)

were employed in the study. ALOS-2 PALSAR-2 level 2.1 data with 10 m resolution including HH and HV polarizations and Sentinel-1 data with 10 m resolution and VH and VV polarizations were used in the present study. To calculate the NDVI, optical satellite data from the Landsat 8 OLI and Sentinel-2 MSI were employed. The description of the satellite image used in this study is provided in *Table 2*.

Table 2. *Satellite imagery used in this study*

No.	Satellite Sensors	Observation date (day, month, year)	Pixel resolution (m)	Polarizations/Band (Indices used)
1	ALOS-2 PALSAR-2	30-08-2018	10	HH, HV
2	Sentinel-1	02-11-2018	10	VH, VV
3	Landsat 8 OLI	31-10-2018	30, 15	Band 4, Band 5, Band 8 (NDVI)
4	Sentinel-2 MSI	02-11-2018	10	Band 4, Band 8 (NDVI)

Satellite data processing and variable extraction

The RADAR images were radiometrically corrected, filter to remove noise, and terrain corrected.

Digital number (DN) values from ALOS-2 PALSAR-2 Level 2.1 imagery were converted into sigma naught backscattering intensities (σ^0) in *Equation 2* (JAXA, 2017).

$$\sigma^0 = 10 \times \log_{10}(DN^2) + CF \quad (\text{Eq.2})$$

where, σ^0 is backscattering intensity (dB), DN is digital number values, and CF is the calibration factor (constant, set at -83 (JAXA, 2017)).

Similarly, the DN values from Sentinel-1 data were converted to the backscatter values using *Equation 3* (ESA, 2019).

$$\sigma^0 = DN^2 / A_{\sigma}^2 \quad (\text{Eq.3})$$

where, σ^0 is backscattering intensity (dB), DN is digital number values and A_{σ}^2 is the reference area of the scattering surface.

Image texture was analyzed using the Gray Level Co-occurrence Matrix (GLCM) technique applied to different polarizations of ALOS-2 PALSAR-2 data. The extracted texture variables included contrast, correlation, dissimilarity, entropy, homogeneity, mean, second angular moment and variance (Haralick et al., 1973).

Top-of-atmosphere (TOA) reflectance data from Landsat 8 OLI and Sentinel-2 were acquired and pre-processed, focusing on cloud-free observations over the study area. The NDVI was then calculated for both sensors using *Equation 4* (Rouse et al., 1974).

$$NDVI = \frac{NIR-RED}{NIR+RED} \quad (\text{Eq.4})$$

where NIR is the near-infrared spectral band and RED is the red spectral band.

The sample plot positions were used to extract the independent variables for AGB model development.

Cross-validation approach

In this study, the model training was conducted using a k-fold cross-validation approach. The k-fold approach works by dividing the training set into k subsets of equal size without a common element. In each run, one of the 'k' subsets is taken out as a validation set and the model is built using the remaining 'k-1' subsets. The final model is determined based on the average of the training and validation errors. In this study, the training data was split into 4 subsets ($k = 4$).

Model validation and accuracy assessment

A set of 25% of the sample plots was selected randomly for validation purposes. This dataset is independent of the dataset used to run the model. Model accuracy was assessed using the coefficient of determination (R^2 ; Equation 5) and the root mean square error (RMSE; Equation 6) as performance evaluation metrics.

$$R^2 = 1 - \frac{\text{Explained Sum of Squares}}{\text{Total Sum of Squares}} \quad (\text{Eq.5})$$

$$\text{RMSE} = \sqrt{\frac{1}{n} \sum_{j=1}^n (\text{Predicted AGB value} - \text{Observed AGB value})^2} \quad (\text{Eq.6})$$

where the predicted AGB value is the AGB value obtained from the model; Observed AGB biomass value is the AGB value from the inventory data, and n is the number of sample plots used.

Result

AGB models using parameters from a single satellite

Using the variables extracted from single satellite data, various AGB models were constructed using the k-fold cross-validation technique. A total of 20 models, using the single satellite variables were constructed, and their cross-validation R^2 and RMSE were calculated (Table 3).

Table 3 shows that some of these models (model 4, model 18, and model 20) showed high R^2 /RMSE values (0.83/9.88, 0.82/10.17, and 0.82/10.17).

Biomass models using a combination of parameters from different satellites

AGB models were developed by integrating variables derived from ALOS-2 PALSAR-2, Landsat 8 OLI, and Sentinel-2 datasets. A total of 10 combined models were constructed (Table 4). The best model ($R^2 = 0.84$, $\text{RMSE} = 9.59$) was achieved using a combination of variables, including HV polarization backscatter and GLCM-derived texture features from ALOS-2 PALSAR-2, along with NDVI from Sentinel-2 imagery.

Biomass mapping

The model 28 consisted of ALOS-2 PALSAR-2 HV polarization, GLCM from HV polarization, and Sentinel-2 NDVI as independent variables. On validation, the model was found to have an R^2 of 0.84, $\text{RMSE} = 9.59 \text{ Mg ha}^{-1}$ (Figure 3). Model 28, which

demonstrated the highest predictive performance, was applied to map the spatial distribution of AGB in the study area (Figure 4).

Based upon the forest AGB map, it was found that of the total area of 8,038.26 ha, 16.01% (1,286.92 ha) is rich forest biomass area with forest AGB ranging from 100 Mg ha⁻¹ and above. Whereas 39.75% (3,195.11 ha) of the area is medium biomass forest with AGB ranging between 50 to 100 Mg ha⁻¹, and 32.63% (2,623.07 ha) of the area was low biomass forest with AGB lower than 50 Mg ha⁻¹. and 11.61% (933.16 ha) of non-forest.

Table 3. Summary of the AGB estimation models

No.	Variable used	AGB model	R ²	RMSE
ALOS-2 PALSAR-2 satellite (L-band)				
Model 1	HH	AGB = 92.54 + 8.009*HH	0.34	19.47
Model 2	HH	AGB = 19.14 + 87.74*HH + 59.55*HH ²	0.50	16.95
Model 3	HV	AGB = 199.13 + 11.97*HV	0.68	13.56
Model 4	HV	AGB = 19.14 + 123.67*HV + 62.76*HV ²	0.83	9.88
Model 5	(HH+HV)/2	AGB = 161.27 + 11.74*(HH+HV)/2	0.59	15.34
Model 6	(HH+HV)/2	AGB = 19.15 + 114.67*(HH+HV)/2 + 66.73*[(HH+HV)/2] ²	0.79	10.98
Model 7	(HH×HH/HV) + HV	AGB = 140.07 + 5.83*(HV+HH*HH/HV)	0.51	16.78
Model 8	(HH×HH/HV) + HV	AGB = 19.14 + 106.57*(HV+HH*HH/HV) + 69.63*(HV+HH*HH/HV) ²	0.72	12.68
Model 9	HH, HV	AGB = 198.14 + 0.57*HH + 11.55*HV	0.68	13.56
Model 10	Grey-level co-occurrence matrix (GLCM)	AGB = -143.21 - 5.17*Contrast + 1.68*Correlation + 109.61*Dissimilarity - 71.14*Entropy - 517.48*Homogeneity + 0.94*Mean - 794.70*SecondMoment - 2.13*Variance	0.24	20.89
Model 11	GLCM and HH	AGB = -409.99 + 11.50*HH - 1.54*Contrast + 27.32*Correlation + 60.72*Dissimilarity + 148.43*Entropy + 333.65*Homogeneity - 6.45*Mean + 2132.99*SecondMoment - 2.43*Variance	0.65	14.18
Model 12	GLCM and HV	AGB = 109.07 + 12.27*HV - 2.81*Contrast + 4.27*Correlation + 40.76*Dissimilarity - 41.54*Entropy + 228.71*Homogeneity + 1.64*Mean - 192.38*SecondMoment + 2.78*Variance	0.79	10.98
Sentinel-1 satellite (C-band)				
Model 13	VH	AGB = 145.06 + 8.08*VH	0.36	19.17
Model 14	VH	AGB = 19.14 + 90.38*VH + 51.11*(VH) ²	0.48	17.28
Model 15	VV	AGB = 88.58 + 5.95*VV	0.24	20.89
Model 16	VV	AGB = 19.14 + 73.74*VV + 46.85*(VV) ²	0.34	19.47
		AGB = 12.58 + 12.41*VH - 4.28*VV	0.39	18.72
Landsat 8 OLI satellite				
Model 17	Landsat 8 Normalized difference vegetation index (NDVI_LS8)	AGB = -42.45 + 130.23*NDVI_LS8	0.61	14.97
Model 18	NDVI_LS8	AGB = 19.14 + 117.14*NDVI_LS8 + 67.41*(NDVI_LS8) ²	0.82	10.17
Sentinel-2 satellite				
Model 19	Sentinel 2 Normalized difference vegetation index (NDVI_S2)	AGB = - 45.68 + 140.58*NDVI_S2	0.68	13.56
Model 20	NDVI_S2	AGB = 19.15 + 123.42*NDVI_S2 + 56.42*(NDVI_S2) ²	0.82	10.17

Table 4. Summary of the AGB estimation models developed using integrated optical and radar satellite data

No.	Variable used	AGB model	R ²	RMSE
ALOS – 2 PALSAR -2 and Landsat 8 OLI satellites				
Model 21	GLCM and NDVI_LS8	AGB = 89.82 - 2.25*Contrast + 4.53*Correlation + 39.90*Dissimilarity - 77.74*Entropy + 201.60*Homogeneity + 1.08*Mean - 1460.66*SeconMoment - 0.55*Variance + 122.30*NDVI_LS8	0.65	14.18
Model 22	HH, GLCM and NDVI_LS8,	AGB = - 164.02 + 7.35*HH + 78.50*NDVI_LS8 - 0.97*Contrast + 19.89*Correlation + 33.63*Dissimilarity + 64.90*Entropy + 197.27*Homogeneity - 3.68*Mean + 648.33*SeconMoment - 1.31*Variance	0.76	11.74
Model 23	HV, GLCM and NDVI_LS8	AGB = 137.04 + 9.63*HV + 43.22*NDVI_LS8 - 2.293*Contrast + 4.72*Correlation + 30.96*Dissimilarity - 50.25*Entropy + 179.33*Homogeneity + 1.54*Mean - 557.59*SecondMoment + 2.28*Variance	0.82	10.17
Model 24	HH, NDVI_LS8	AGB = 6.31 + 4.19*HH + 108.31*NDVI_LS8	0.69	13.34
Model 25	HV, NDVI_LS8	AGB = 106.77 + 7.91*HV + 66.22*NDVI_LS8	0.76	11.74
ALOS – 2 PALSAR -2 and Sentinel -2 satellites				
Model 26	NDVI_S2, GLCM	AGB = - 1.92 + 132.08*NDVI_S2 - 1.53*Contrast - 6.84*Correlation + 27.23*Dissimilarity - 38.00*Entropy + 170.15*Homogeneity + 0.56*Mean - 1071.15*SeconMoment + 0.13*Variance	0.72	12.8
Model 27	HH, GLCM and NDVI_S2	AGB = - 191.59 + 6.37*HH + 92.82*NDVI_S2 - 0.60*Contrast + 9.88*Correlation + 24.65*Dissimilarity + 73.72*Entropy + 171.60*Homogeneity - 3.41*Mean + 632.09*SeconMoment - 0.70*Variance	0.80	10.72
Model 28	HV, GLCM and NDVI_S2	AGB = 94.90 + 8.35*HV + 61.99*NDVI_S2 - 1.8637*Contrast - 0.55*Correlation + 24.06*Dissimilarity - 35.43*Entropy + 157.81*Homogeneity + 1.24*Mean - 514.31*SeconMoment + 2.27*Variance	0.84	9.59
Model 29	HH, NDVI_S2	AGB = -5.09 + 3.40*HH + 120.28*NDVI_S2	0.73	12.45
Model 30	HV, NDVI_S2	AGB = 86.74 + 6.98*HV + 81.17*NDVI_S2	0.79	10.98

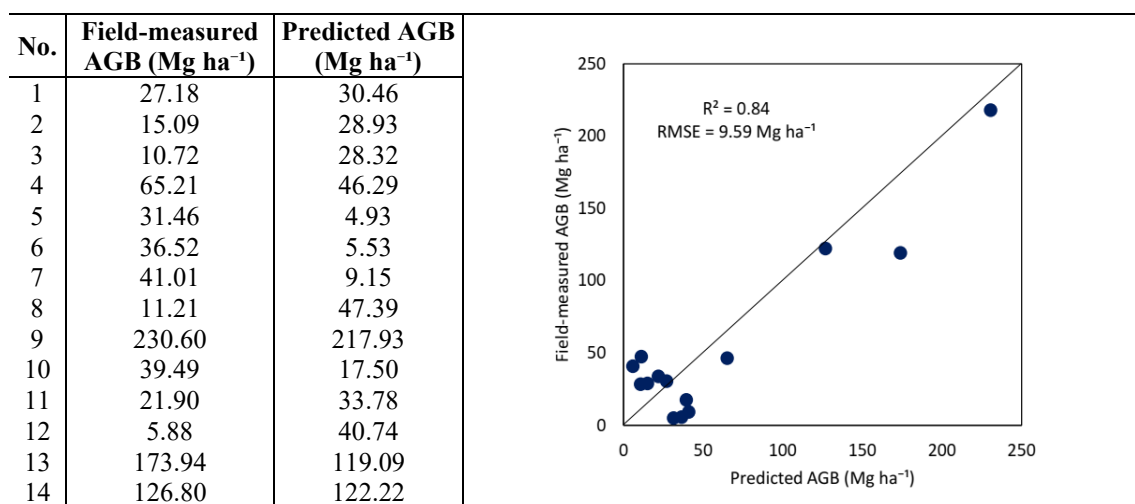


Figure 3. The validation results are shown in the 1:1 cross-plot comparing predicted and field-measured AGB

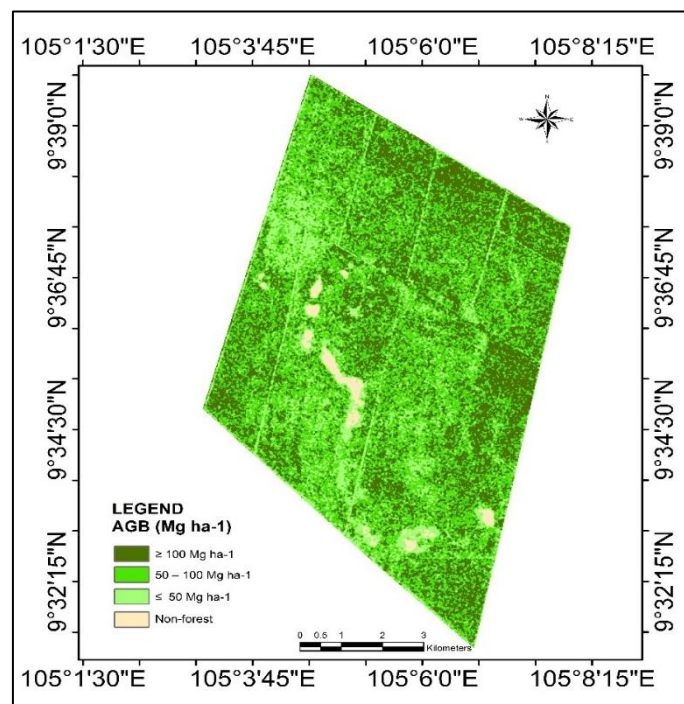


Figure 4. Spatial distribution of forest AGB in U Minh Thuong National Park

Discussions

In this study, satellite data from ALOS-2 PALSAR-2, Sentinel-1, Sentinel-2 and Landsat 8 OLI satellites were utilized to assess the applicability of variables extracted from these datasets, such as backscattering intensity, GLCM-based textures, and NDVI for estimating *Melaleuca* forest AGB. The variables extracted from these satellite datasets were found to be highly correlated with the field-measured AGB in the *Melaleuca* forest, except for data from the Sentinel-1 satellite, because this limitation also applies to the C-band SAR in moist tropical forests due to poor canopy penetration. Despite this exception, our results surpass those of previous studies conducted by Omar et al. (2020), Khan et al. (2020) and Stratoulis et al. (2022) when using only one type of optical satellite imagery.

Our analysis reveals that each satellite, including ALOS-2 PALSAR-2, Landsat 8 OLI, and Sentinel-2, has the potential to estimate the forest AGB with high accuracy. Each of these datasets showed high R^2 for at least one of the extracted variables while developing the forest AGB models. The performances of these simple models can also be attributed to the fact that the AGB of the *Melaleuca* forest in the study area is not extremely high. However, the combination of multiple satellite data sensors enhanced the performance of the biomass model. This integration has reduced the problem of data saturation, and has improved the estimation of forest AGB (Zhao et al., 2016; Khan et al., 2020; Ahmad et al., 2021).

When considering only the variables extracted from individual satellite datasets as the independent variables for constructing the biomass models, Model 4, utilizing the backscatter values from HV polarization of ALOS-2 PALSAR-2 data, Model 18 utilizing the NDVI calculated from Landsat 8 OLI data, and Model 20 utilizing the NDVI calculated from Sentinel-2 data exhibited noteworthy performance, explaining 84%, 82%, and 82% variability in the *Melaleuca* forest AGB, respectively.

Models developed using the combination of parameters from different satellites yielded high R^2 values. Model 23 incorporating ALOS-2 PALSAR-2 HV data, GLCM textures, and Landsat 8 NDVI ($R^2 = 0.82$; $RMSE = 10.17 \text{ Mg ha}^{-1}$), Model 27 integrating ALOS-2 PALSAR-2 HH data, textures, and Sentinel-2 NDVI ($R^2 = 0.80$; $RMSE = 10.72 \text{ Mg ha}^{-1}$), and Model 28 combining ALOS-2 PALSAR-2 HV data, textures, and Sentinel-2 NDVI ($R^2 = 0.84$; $RMSE = 9.59 \text{ Mg ha}^{-1}$) exhibited high accuracies.

Our results surpass those of recent studies, such as those by Vafaei et al. (2018), Zhu et al. (2020), and López et al. (2020), due to the unique characteristics of our study area, characterized by flat topography, sparse forest tree density, predominantly small to medium-sized trees, and relatively low biomass reserves. Consistent with prior research by Lucas et al. (2010) and Luong et al. (2016, 2019), we also found that backscattering intensity at HV polarization exhibits greater sensitivity compared to HH polarization. This can be attributed to the fact that the dominant scattering mechanisms in forests, such as volume scattering, are more effectively captured by HV polarization due to its ability to penetrate vegetation canopies and interact with vertical components of vegetation, such as stems and branches.

Conclusions

This study utilized ALOS-2 PALSAR-2, Sentinel-1, Sentinel-2, and Landsat 8 OLI satellites to estimate the spatial distribution of AGB in the *Melaleuca* forest in the Mekong Delta region of Vietnam. In total, 30 models using various variables from individual satellite datasets and combinations of these datasets were constructed using a k-fold cross-validation approach. The results reflected that in the present ecological settings, where *Melaleuca* forest with moderate forest AGB grows on peat soils, the capacity of RADAR (ALOS-2 PALSAR-2) and optical (Landsat 8, Sentinel-2) datasets for AGB estimation was comparable. This study reiterates the fact that the sensitivity of backscattering intensity at HV is more than that of backscattering intensity at HH. The study also highlighted the importance of the synergistic use of RADAR and optical data for forest AGB estimation. The results of this study offer a scientific foundation to support decision-making related to conservation, sustainable development strategies, and potential engagement in carbon emission trading mechanisms for U Minh Thuong National Park, located in the Mekong Delta region of Vietnam.

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